



HAL
open science

Monitoring microplastics in the Seine River in the Greater Paris area

Cleo Stratmann, Rachid Dris, Johnny Gasperi, Frans Buschman, Adriaan Markus, Sabrina Guerin, A. Dick Vethaak, Bruno Tassin

► **To cite this version:**

Cleo Stratmann, Rachid Dris, Johnny Gasperi, Frans Buschman, Adriaan Markus, et al.. Monitoring microplastics in the Seine River in the Greater Paris area. *Frontiers in Earth Science*, 2024, 12, 10.3389/feart.2024.1386547 . hal-04609944

HAL Id: hal-04609944

<https://univ-eiffel.hal.science/hal-04609944>

Submitted on 17 Jun 2024

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution 4.0 International License



OPEN ACCESS

EDITED BY

Uwe Schneidewind,
University of Birmingham, United Kingdom

REVIEWED BY

Reza Dehbandi,
University of Birmingham, United Kingdom
Mohammad Wazne,
Université Claude Bernard Lyon 1, France

*CORRESPONDENCE

Cleo N. Stratmann,
✉ c.stratmann@posteo.de
Rachid Dris,
✉ rachid.dris@u-pec.fr

†Deceased

RECEIVED 15 February 2024

ACCEPTED 07 May 2024

PUBLISHED 12 June 2024

CITATION

Stratmann CN, Dris R, Gasperi J,
Buschman FA, Markus AA, Guerin S,
Vethaak AD and Tassin B (2024), Monitoring
microplastics in the Seine River in the Greater
Paris area.
Front. Earth Sci. 12:1386547.
doi: 10.3389/feart.2024.1386547

COPYRIGHT

© 2024 Stratmann, Dris, Gasperi, Buschman,
Markus, Guerin, Vethaak and Tassin. This is an
open-access article distributed under the
terms of the [Creative Commons Attribution
License \(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or
reproduction in other forums is permitted,
provided the original author(s) and the
copyright owner(s) are credited and that the
original publication in this journal is cited, in
accordance with accepted academic practice.
No use, distribution or reproduction is
permitted which does not comply with
these terms.

Monitoring microplastics in the Seine River in the Greater Paris area

Cleo N. Stratmann^{1*}, Rachid Dris^{1*}, Johnny Gasperi²,
Frans A. Buschman³, Adriaan A. Markus³, Sabrina Guerin⁴,
A. Dick Vethaak^{3,5†} and Bruno Tassin¹

¹Laboratoire Eau Environnement et Systèmes Urbains, Ecole des Ponts ParisTech, Université Paris Est Créteil, Marne-la-Vallée, France, ²Laboratoire Eau et Environnement GERS-LEE, Université Gustave Eiffel, Bouguenais, France, ³Deltares, Delft, Netherlands, ⁴Syndicat Interdépartemental pour l'Assainissement de l'Agglomération Parisienne (SIAAP), Paris, France, ⁵Amsterdam Institute for Life and Environment (A-LIFE), Vrije Universiteit Amsterdam, Amsterdam, Netherlands

This data paper presents microplastic (MP) concentration and composition data from four campaigns in a one-year period in the Seine River within the Greater Paris area. Detailed data on MP occurrence, physical water parameters, hydrological conditions, and potential sewer point sources are presented. Initial analysis of the data set shows that MP concentrations were significant (median: 600 particles m⁻³ and 100 mg m⁻³) and MP fluxes showed variations across sampling sites and campaigns. Three polymer types (polypropylene - PP, polyethylene - PE, polystyrene - PS) account for 96% of the total number of MP. This dataset provides valuable insights for understanding river MP pollution under anthropogenic pressure.

KEYWORDS

microplastic fluxes, water balance, river catchment, urban plastic pollution, sewer discharges

1 Introduction

Microplastics (MPs) are ubiquitous environmental contaminants present in all natural compartments (Miranda et al., 2020). MPs are persistent and harmful, e.g., through the release of toxic chemicals (Wagner et al., 2024), negatively impact natural resources, organisms, and human health (Amobonye et al., 2021). It is thus important to understand MPs fate and impact once they leaked into the environment to adopt mitigation and protection measures. Freshwaters are key ecosystems worthy of protection against MP pollution and rivers play an important role. Rivers are often under anthropogenic pressure receiving MPs (Kumar et al., 2021) through surface runoff, combined sewer overflows (CSOs), or wastewater treatment plant (WWTP) effluent (Fahrenfeld et al., 2019; Nguyen et al., 2024). To elucidate sources, understand MP transport and assess the ecological impacts comprehensive monitoring is needed. This should include MP concentration and fluxes (Miranda et al., 2020) across temporal and spatial scales like in a recent study on microplastics in canals of Amsterdam (Sefiloglu et al., 2024). Additional information on the environmental conditions such as the river flow rate, weather conditions, and water quality parameters like suspended particle concentration or algal biomass help to explain MPs in the environmental context (Birch et al., 2020) and to identify potential solutions (Li et al., 2020; Li et al., 2023).

The Seine River in France (Figure 1) meanders 750 km from east to northwest and is subject to intense urban activities (up to 5,000 inhabitants km⁻²) before reaching the English Channel. Traversing Greater Paris with 12 million inhabitants (Tabuchi et al., 2016) the Seine consistently receives sewer effluents via WWTPs (2021: ~853 million m³ from four major WWTPs) and CSOs (~23 million m³ in 2021) during periods of intense rainfall (Flipo et al., 2020, 2021). Studies (Dris et al., 2015; Treilles et al., 2022) have highlighted MP contamination in the Seine around Paris, exhibiting 4–5,000 particles m⁻³ as median concentration.

Further insights into the complex dynamics of MP occurrence in the Seine River are needed. This data paper is the basis for further analyses presented later answering questions like: Are MP concentrations increased downstream of Paris, indicating significant urban contribution to the contamination and which urban sources could be important contributors? How is MP occurrence linked to the different MP sources along the river stream?

Therefore, a 1-year MP monitoring campaign in the Seine River upstream and downstream of Greater Paris was conducted. This paper describes the unique MP monitoring dataset of the Seine River including 1) an overview of MP occurrence in the Seine between July 2021 and July 2022 based on four sampling campaigns, 2) data on physical water parameters, 3) data on hydrological conditions including a water balance, and 4) data on potential sewer point sources. This dataset provides comprehensive data reporting to foster cross-study comparisons (Cowger et al., 2020).

2 Methods

Monitoring included collecting MP data and some physical water quality and hydrological parameters from field samples and measurements. Detailed hydrological and meteorological data (river discharge, water level, precipitation) and data on sewage discharges (WWTPs, CSOs) were collected by third parties.

2.1 Monitoring

2.1.1 Sampling sites and monitoring period

The monitoring area (Figure 1A) is situated in the upper Seine catchment (~67 × 10³ km²). The oceanic climate has annual average temperatures of 19.5°C in summer and 5°C in winter, with low temporal variations in rainfall (mean annual precipitation 642 mm). The Seine (mean discharge in Paris: 319 m³s⁻¹) has two distinct seasonal flow regimes: a low flow during summer (mean 125 m³s⁻¹) and a high flow during winter (mean 583 m³s⁻¹) (Flipo et al., 2021). This seasonality is driven by evapotranspiration in the catchment and the presence of natural aquifers. Four upstream reservoirs maintain summer flows above 100 m³s⁻¹ and contribute to regulate flood periods. Twelve tributaries flow into the Seine within the monitored area, including the three preeminent tributaries Yonne, Marne, and Oise. Six sampling sites reached downstream spanning 442 km along the Seine from Marnay-sur-Seine to Poses before the estuary influence (Figure 1). The two sampling sites Marnay-sur-Seine (S1) and Choisy-le-Roi (S2) are located upstream of Paris,

and four sampling sites at the locations Suresnes (S3), Bougival (S4), Triel-sur-Seine (S5), and Poses (S6) are located downstream of Paris. An additional sampling site just before the Seine-Marne confluence at Marne à Alfortville (M1) was monitored to account for MPs entering from the Marne River. All sampling sites exhibited anthropogenically modified riverbanks varying from concrete walls to overgrown slopes. Two sampling campaigns per river flow season (low and high) were carried out. The four sampling campaigns were conducted in July 12–22, 2021, November 2–11, 2021, February 15–24, 2022, and July 20–27, 2022.

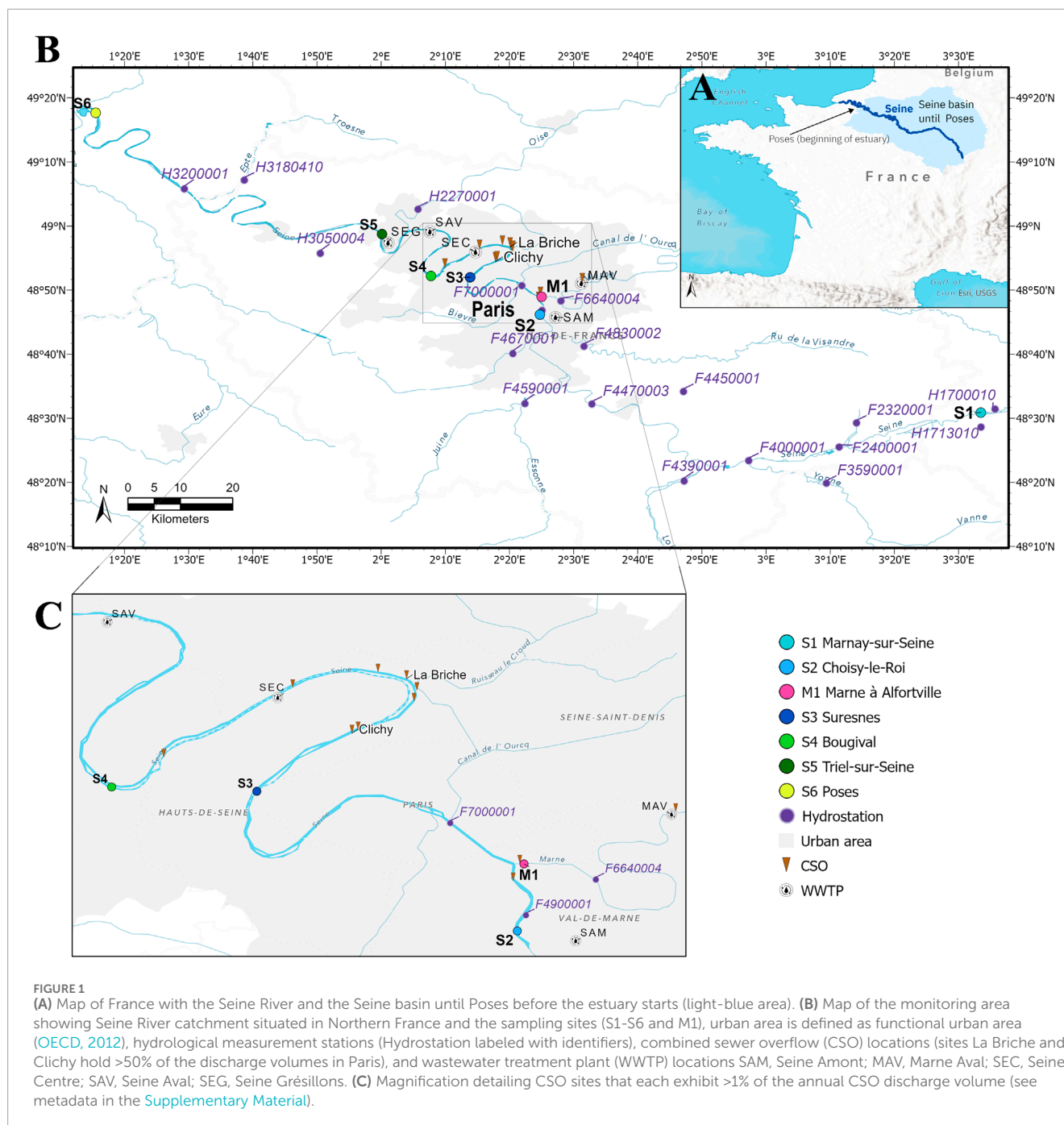
2.1.2 Sample and data collection

2.1.2.1 Microplastics

MPs with a major particle diameter between 25 µm and 300 µm were assessed. The lower sampling mesh size was 10 µm. MP samples were collected ca. 1–3 m away from the river shore at the sampling sites during daylight for all four sampling campaigns in the surface water (upper 0.1–0.3 m) using an *in-situ* cascade filtration pump [Universal Filtration Object (UFO) developed by Aalborg University, (Rist et al., 2020), Supplementary Figure S1]. The same sampling approach was used by Sefiloglu et al. (2024) to determine microplastic concentrations in Amsterdam surface waters. Seven liters per minute of water were pumped through a 5,000 µm grid cage and stainless-steel tube over stainless steel filters of first 300 µm and then 10 µm pore-size. One sample consisted of four filters (16 cm in diameter and 10 µm pore size). A pressure threshold of 1.8 bar indicated filter clogging. Sampling volumes vary from 74L to 940L depending on the water conditions as higher suspended particulate matter in the water led to faster clogging.

The loaded filters were rinsed and ultrasonicated (2–4 min) before hydrogen peroxide (10 vol% H₂O₂) wet oxidation for 18–24 h at 30°C, and the solution was filtered afterwards. This was followed by sodium-iodide (density 1.63–1.7 g cm⁻³ NaI) density separation using JAMSS density separator units (Nakajima et al., 2019) and volume-reducing “anodisc” (0.2 µm, Whatman) filtration. During the procedures, the filters were stored in clean glass Petri dishes at room temperature. After oxidation and density separation, the filters were rinsed and ultrasonicated. The extracted particles on the anodisc filters were then analyzed by micro-Fourier transform infrared spectroscopy (µ-FTIR, Supplementary Material S2) with a detection limit (pixel resolution) of 25 µm. Spectra were further interpreted using systematic identification of MP in the environment (siMPle[®]) software [Supplementary Material S3, Primpke et al. (2020)] providing characteristics such as polymer type, particle dimensions, and mass (calculated by the software using minor and major dimension, particle thickness estimated at 0.6 × 2 × minor dimension, polymer density, and assuming an ellipsoid shape). The sample processing is visualized in Supplementary Material S4.

Throughout the study, strict protocols were followed to ensure the integrity of the samples and prevent MP contamination. To minimize the risk of synthetic material contamination, all personnel always wore cotton lab coats and refrained from wearing synthetic clothing. All procedures were executed under a laminar flow bench or fume hood. All solutions used were pre-filtered (GF/D 2.7µm, Whatman). Glassware, glass-fiber, and stainless-steel filters were muffled, i.e., temperature treatment at 500°C for 2–3 h. Workspaces



were routinely cleaned. Plastic materials were avoided during laboratory processes except for sample processing for campaigns July 2021 and November 2021, where polyethylene (PE) squeezing bottles with polypropylene (PP) caps were utilized for rinsing. Procedural blanks (~100 mL of filtered tap water) were carried out in parallel to the sample processing as of the H_2O_2 oxidation for MPs (results see [Supplementary Material](#)).

2.1.2.2 Physical water quality parameters

Temperature, pH, and conductivity were measured during each MP sampling campaign using a multiprobe

(Multiline P4, WTW). A clean stainless-steel bucket was filled with >5L of river surface water, and measurements were taken inside the bucket. Turbidity was measured as nephelometric turbidity units (NTU) in triplicate using an *in-situ* turbidity meter (Hach 2100P Turbidimeter). Suspended sediment concentration (SSC) was determined from a 1-L sample of river water following the standard procedure ASTM D3977 with a slight modification of drying temperature (>48 h at 60°C). The total sample was filtered on a muffled GF/F filter (pore size 0.7 μ m, Whatman), dried and weighted.

2.1.2.3 Hydrological data

For each sampling site, the distance from the river source and the river catchment area up to that point are obtained from HydroSHEDS data (Lehner and Grill, 2013).

Data of river discharge (Q , m^3s^{-1}) and water level (H , m) for the Seine and Marne Rivers and relevant tributaries near the confluences were obtained from 18 hydrological monitoring stations as daily averaged values from the Central Service for Hydrometeorology and Support for Flood Forecasting in France (Hydro Eaufrance, 2023).

Flow velocity (five replicates, $m s^{-1}$) was measured in the surface water during MP sampling with a portable flowmeter (Flo-mate Model 2000; Marsh-McBirney Inc.). Measurements were not always possible or reliable due to challenging conditions, especially low water flow close to the shore, and the limitation of the instruments' accuracy [$\pm 2\%$ plus zero stability ($1.5 m s^{-1}$)].

Daily precipitation data were obtained from Prevision-Meteo (2023) for three weather stations for the months of the sampling campaigns. Data from station Paris-Montsouris, located in central Paris, were used for sampling sites M1, S2, S3, S4, and S5. Precipitation data from Paris Melun-Villaroche and Evreux-Fauville were related to sampling sites S1 and S6, respectively.

2.1.2.4 Data on sewage discharges as microplastic sources

The acquired data include locations and daily discharge volumes for five selected WWTP effluent locations and twelve of 38 CSO outfall locations, each holding $>1\%$ of discharge and together presenting 91% of the CSO volumes discharged during the monitoring period (Figures 1B, C). Data were provided by SIAAP. The CSO outfall sites Clichy and La Briche are located downstream of sampling site S3, holding 31% and 25%, respectively.

Four large WWTPs are located along the Seine in the Paris urban area. Notably, the WWTP Paris Seine-Amont (SAM) is located a few kilometers before sampling site S2, while the WWTP Seine-Centre (SEC) is upstream of site S4. WWTPs Seine-Aval (SAV) and Seine Grésillons (SEG) are between S4 and S5. WWTP Marne Aval (MAV) is in the Marne River and related to sampling site M1.

2.2 Data analysis

We used the free software for statistical computing R (version 4.3.0) to conduct the data analyses (R Core Team, 2021). A Non-parametric Spearman rank test was applied to assess correlations between variables. Replicates of physical water parameters were averaged. For *MP summary* (Supplementary Material) MP (mass) concentrations and fluxes were reported as a sum of all individual particles per sample. MP polymer type proportions and sizes per campaign were weighted by sample size.

2.2.1 Microplastic data

MP data exploration followed the protocol outlined by Zuur et al. (2010). Some particles with major dimension above $300 \mu m$ (6.2%) went through the filter during sampling. For comparability, MP particles from 25 to $300 \mu m$ were analyzed. The size limits were chosen to focus on smaller MP with comparability to other studies (Dris et al., 2024).

MP numeric (further on referred to as MP concentration) and mass concentrations (Eqs 1, 2) were calculated as the number of observations (N) or particle mass divided by the sampling volume (V_s) of the respective sample and converted into number of particles m^{-3} and $\mu g L^{-1}$. The concentrations were estimated for different groups, e.g., per campaign and sampling site, and polymer type.

$$MP \text{ concentration} [particles m^{-3}] = \frac{N [particles]}{V_s [m^3]} \quad (1)$$

$$MP \text{ mass concentration} [\mu g L^{-1}] = \frac{mass}{V_s [L]} \quad (2)$$

The MP (mass) flux (Eqs 3, 4) was calculated as MP (mass) concentration multiplied by the river discharge (Q) per sampling event as particles s^{-1} or $mg s^{-1}$.

$$MP \text{ flux} [particles s^{-1}] = MP \text{ concentration} [particles m^{-3}] * Q [m^3 s^{-1}] \quad (3)$$

$$MP \text{ mass flux} [mgs^{-1}] = MP \text{ mass concentration} [mg m^{-3}] * Q [m^3 s^{-1}] \quad (4)$$

2.2.2 Water balance

For the 1-year monitoring period (with sampling durations of about 2 weeks per campaign) a water balance (Supplementary Material S5) of the Seine River discharge was conducted to quantify the hydrodynamic conditions and to study MP transport and fate. Balances were generated for the hydrological stations in the Seine. Discharge data of the Seine and tributaries were considered. The balances were calculated as the differences between the inflows and outflows. The downstream discharge Q_x (Eq. 5) is the sum of the previous upstream discharge (Q_0) and the incoming tributary discharges (Q_i) between Q_0 and Q_x . Acknowledging that ground water flow into the river and evaporation are to be assumed negligible with respect to the discharge, the difference (Eq. 6) should be close to $0 m^3 s^{-1}$.

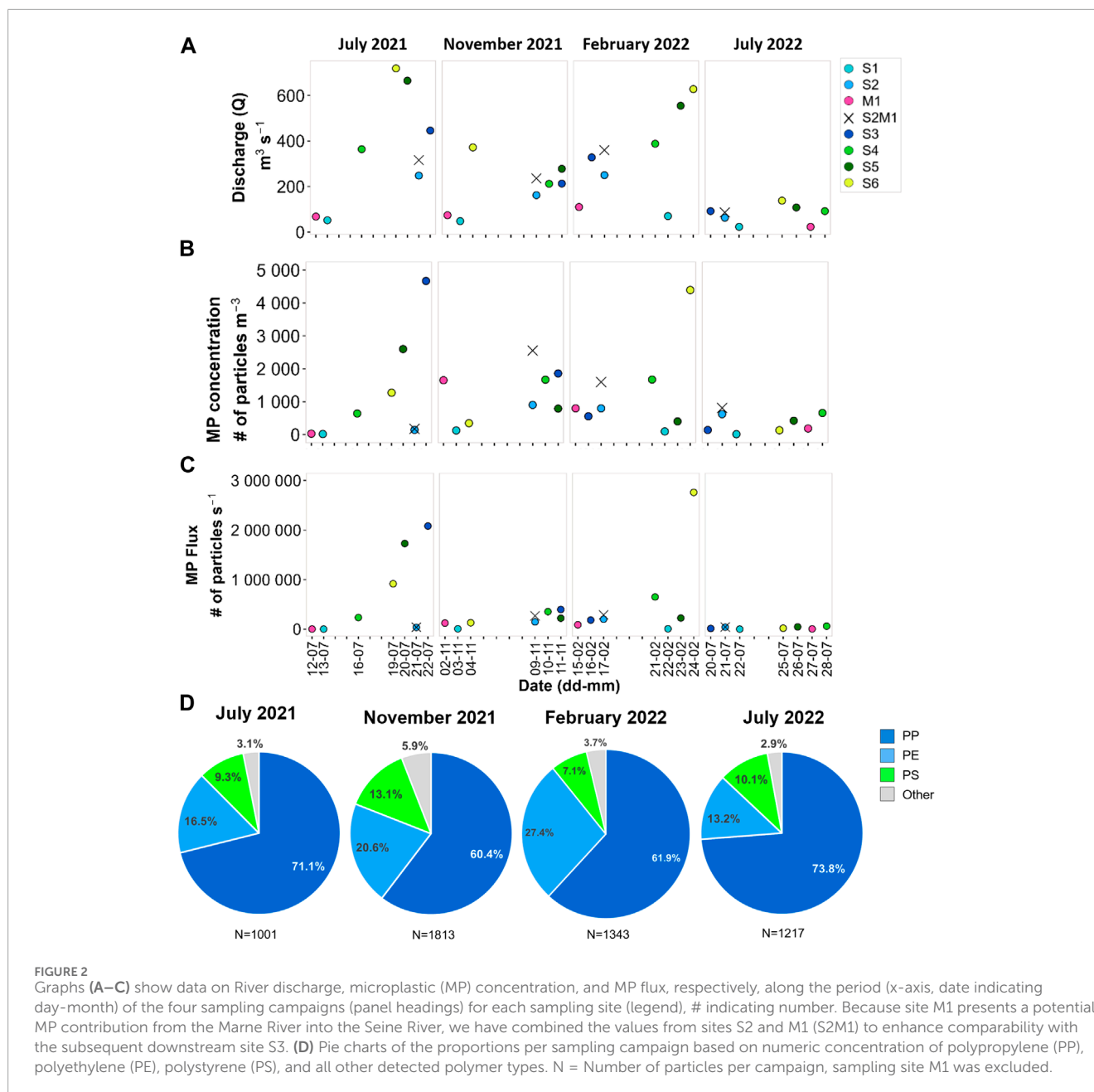
$$Q_x = Q_0 + \sum Q_i \quad (5)$$

$$Difference = Q_x - Q_0 - \sum Q_i \stackrel{def}{=} 0 \quad (6)$$

3 Description of the data and initial analyses

3.1 Microplastics

The dataset contains comprehensive data of individual MP particles (total $N = 5,922$) per sampling site and sampling day, detailing particle characteristics (mass, polymer type, minor, major, and ferret dimension). For number and mass, MP concentrations and MP fluxes were estimated for each campaign per sampling date, and sampling site (Figures 2B, C; Supplementary Material MP summary). The median MP concentration was $\sim 600 particles m^{-3}$ and MP flux $165 \times 10^3 particles s^{-1}$. MP mass concentrations and



mass fluxes ranged between 2 and 960 $mg m^{-3}$, and 57 and 500 $\times 10^3 mg s^{-1}$, respectively. The annual estimated MP concentration and MP mass flux in the Seine are 5.19×10^{12} particles yr^{-1} and 816 $t yr^{-1}$, respectively.

MP concentrations varied across sampling sites and campaigns between 14 particles m^{-3} (S1) and 4,700 particles m^{-3} (S3). MP fluxes ranged from 300 particles s^{-1} (S1) to 2.67×10^6 particles s^{-1} (S6). For each campaign, the lowest MP concentrations were always found at the upstream sampling site S1. The MP fluxes for S1 (ranging from 300 to 6,000 particles s^{-1}) are smaller than the MP flux values of sampling sites S2–S6 (ranging between 13×10^3 and 2.76×10^6 particles s^{-1}), with a median of 165×10^3 particles s^{-1} . MP concentration correlates moderately ($\rho = 0.65$, $p < 0.01$) to

river discharge, MP (mass) flux ($\rho = 0.89$, $p < 0.01$; $\rho = 0.85$, $p < 0.01$) correlates stronger to river discharge.

The MP concentrations in the Seine are comparable to higher concentrations reported in European rivers' surface water (Gao et al., 2023).

Sixteen polymer types were identified including acrylics, epoxy and rubber (Supplementary Material S5), and the most abundant polymer types across all Seine samples were polypropylene (PP) (concentration: 67%, mass concentration: 48%), polyethylene (PE) (19%, 38%), and polystyrene (PS) (10%, 10%), reflecting similar findings in European rivers (Scherer et al., 2020; Gao et al., 2023; Sefiloglu et al., 2024). The polymer type distribution is relatively consistent across the four campaigns (Figure 2D) and

Funding

The author(s) declare that financial support was received for the research, authorship, and/or publication of this article. This project has received funding from the European Union's Horizon 2020 research and innovation programme under grant agreement No. 860720.

Acknowledgments

This study was performed as part of a Ph.D. project in the project LimnoPlast: Microplastics in Europe's Freshwater Ecosystems: from sources to solutions. We would like to express great gratitude to every person who has helped for preparation, in the field, in the laboratory, and for expert input: Mohamed Saad, Philippe Dubois, Azeez Odofin, Daniela Castro, Ayoub Elcadi, Öykü Sefigolu, Guilherme Calabro, Robin Richoux, Erwan Garcia Gonzales, Nadia Bouzid, Gabriël Olthof, Maryem Mehboob, Minh-Trang Nguyen, Robin Treilles, Jean-Sebastien Barbier, Max Beaurepaire, Sinimar Awad, and Céline Hadjebar. We thank the OSU-Efluve for access to the μ -FTIR instrument granted by the PRAMMICS platform. A special thanks goes to PIREN-Seine and MeSeine Innovation SIAAP program for financial support.

References

- Amobonye, A., Bhagwat, P., Raveendran, S., Singh, S., and Pillai, S. (2021). Environmental impacts of microplastics and nanoplastics: a current overview. *Front. Microbiol.* 12, 768297. doi:10.3389/fmicb.2021.768297
- Birch, Q. T., Potter, P. M., Pinto, P. X., Dionysiou, D. D., and Al-Abed, S. R. (2020). Sources, transport, measurement and impact of nano and microplastics in urban watersheds. *Rev. Environ. Sci. Bio/Technology* 19 (2), 275–336. doi:10.1007/s11157-020-09529-x
- Cowger, W., Booth, A. M., Hamilton, B. M., Thaysen, C., Primpke, S., Munno, K., et al. (2020). Reporting guidelines to increase the reproducibility and comparability of research on microplastics. *Appl. Spectrosc.* 74 (9), 1066–1077. doi:10.1177/0003702820930292
- Dris, R., Beaurepaire, M., Bouzid, N., Stratmann, C., Nguyen, M. T., Bordignon, F., et al. (2024). "Chapter 3 - sampling and analyzing microplastics in rivers: what methods are being used after a decade of research?," in *Microplastic contamination in aquatic environments*. Editor E. Y. Zeng Second Edition (Elsevier), 65–91. doi:10.1016/B978-0-443-15332-7.00013-2
- Dris, R., Gasperi, J., Rocher, V., Saad, M., Renault, N., and Tassin, B. (2015). Microplastic contamination in an urban area: a case study in Greater Paris. *Environ. Chem.* 12 (5), 592. doi:10.1071/EN14167
- Fahrenfeld, N. L., Arbuckle-Keil, G., Naderi Beni, N., and Bartelt-Hunt, S. L. (2019). Source tracking microplastics in the freshwater environment. *TrAC Trends Anal. Chem.* 112, 248–254. doi:10.1016/j.trac.2018.11.030
- Flipo, N., Gallois, N., Labarthe, B., Baratelli, F., Viennot, P., Schuete, J., et al. (2020). "Pluri-annual water budget on the Seine basin: past, current and future trends," in *The Seine River basin*. Editors N. Flipo, P. Labadie, and L. Lestel (Springer International Publishing), 59–89. doi:10.1007/978-3-319-392
- Flipo, N., Labadie, P., and Lestel, L. (2021). *The Seine River basin* (Springer International Publishing). doi:10.1007/978-3-030-54260-3
- Gao, S., Orłowski, N., Bopf, F. K., and Breuer, L. (2023). A review on microplastics in major European rivers. *WIREs Water* 11, e1713. doi:10.1002/wat2.1713
- Hydro Eaufrance (2023). HydroPortail. Available at: <https://www.hydro.eaufrance.fr/>.
- Kumar, R., Kumar, R., Sharma, P., Manna, C., Manna, C., Jain, M., et al. (2021). Abundance, interaction, ingestion, ecological concerns, and mitigation policies of microplastic pollution in riverine ecosystem: a review. *Sci. Total Environ.* 782, 146695. doi:10.1016/j.scitotenv.2021.146695
- Lehner, B., and Grill, G. (2013). Global river hydrography and network routing: baseline data and new approaches to study the world's large river systems. *Hydrol. Process.* 27 (15), 2171–2186. doi:10.1002/hyp.9740
- Li, C., Busquets, R., and Campos, L. C. (2020). Assessment of microplastics in freshwater systems: a review. *Sci. Total Environ.* 707, 135578. doi:10.1016/j.scitotenv.2019.135578
- Li, S., Wang, H., He, Y., Liang, D., Shen, Y., Gu, Q., et al. (2023). How microplastic loads relate to natural conditions and anthropogenic activities in the Yangtze river basin. *Chemosphere* 342, 140146. doi:10.1016/j.chemosphere.2023.140146
- Miranda, M. N., Silva, A. M. T., and Pereira, M. F. R. (2020). Microplastics in the environment: a DPSIR analysis with focus on the responses. *Sci. Total Environ.* 718, 134968. doi:10.1016/j.scitotenv.2019.134968
- Nakajima, R., Tsuchiya, M., Lindsay, D. J., Kitahashi, T., Fujikura, K., and Fukushima, T. (2019). A new small device made of glass for separating microplastics from marine and freshwater sediments. *PeerJ* 7, e7915. doi:10.7717/peerj.7915
- Nguyen, M. T., Phuong, N. N., Saad, M., Tassin, B., Gillet, T., Guérin-Rechdaoui, S., et al. (2024). Microplastic accumulation in sewer sediments and its potential entering the environment via combined sewer overflows: a study case in Paris. *Environ. Sci. Pollut. Res.* 31, 10501–10507. doi:10.1007/s11356-023-31734-8
- OECD (2012). Redefining "urban"—a new way to measure metropolitan areas. Available at: <https://www.oecd-ilibrary.org/content/publication/9789264174108-en>.
- Prevision-meteo (2023). Daily climate. Available at: <https://prevision-meteo.ch/>.
- Primpke, S., Cross, R. K., Mintenig, S. M., Simon, M., Vianello, A., Gerds, G., et al. (2020). Toward the systematic identification of microplastics in the environment: evaluation of a new independent software tool (siMPle) for spectroscopic analysis. *Appl. Spectrosc.* 74 (9), 1127–1138. doi:10.1177/0003702820917760
- R Core Team (2021). *R: a language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. Available at: <https://www.R-project.org/>.
- Rist, S., Vianello, A., Winding, M. H. S., Nielsen, T. G., Almeda, R., Torres, R. R., et al. (2020). Quantification of plankton-sized microplastics in a productive coastal Arctic marine ecosystem. *Environ. Pollut.* 266, 115248. doi:10.1016/j.envpol.2020.115248

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/feart.2024.1386547/full#supplementary-material>

- Scherer, C., Weber, A., Stock, F., Vurusic, S., Egerci, H., Kochleus, C., et al. (2020). Comparative assessment of microplastics in water and sediment of a large European river. *Sci. Total Environ.* 738, 139866. doi:10.1016/j.scitotenv.2020.139866
- Sefiloglu, F. Ö., Stratmann, C. N., Brits, M., Van Velzen, M. J. M., Groenewoud, Q., Vethaak, A. D., et al. (2024). Comparative microplastic analysis in urban waters using μ -FTIR and Py-GC-MS: a case study in Amsterdam. *Environ. Poll.* 351, 124088. doi:10.1016/j.envpol.2024.124088
- Tabuchi, J.-P., Tassin, B., and Blatrix, C. (2016). *Grand Paris: Eau et Changement Global*, 24.
- Treilles, R., Gasperi, J., Tramoy, R., Dris, R., Gallard, A., Partibane, C., et al. (2022). Microplastic and microfiber fluxes in the Seine River: flood events versus dry periods. *Sci. Total Environ.* 805, 150123. doi:10.1016/j.scitotenv.2021.150123
- Wagner, M., Monclús, L., Arp, H. P. H., Groh, K. J., Løseth, M. E., Muncke, J., et al. (2024). State of the science on plastic chemicals—identifying and addressing chemicals and polymers of concern. [object Object]. doi:10.5281/ZENODO.10701706
- Zuur, A. F., Ieno, E. N., and Elphick, C. S. (2010). A protocol for data exploration to avoid common statistical problems: *Data exploration. Methods Ecol. Evol.* 1 (1), 3–14. doi:10.1111/j.2041-210X.2009.00001.x